### Comparing qPCR and eDNAmetabarcoding for detection of elusive lampreys

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#### **Abstract**

With the world potentially approaching the sixth mass extinction, it is imperative that species we lack knowledge about are investigated to learn more about their distribution. An efficient way to acquire knowledge about many threatened and elusive freshwater fish species is through environmental DNA monitoring. eDNA monitoring has begun to be a staple surveying tool to investigate distribution of different species due to it being cost- and time-effective, and reliable. There are different methods within eDNA monitoring, such as qPCR and DNA-metabarcoding. qPCR is a species-specific real-time quantitative analysis method, where species are investigated one at a time, whereas DNA-metabarcoding investigate multiple freshwater fish species at once.

In this study I tested and compared qPCR and DNA-metabarcoding methods for studying the distribution of river (Lampetra fluviatilis) and brook (Lampetra planeri) lamprey, in a river system where they are known to reside, Haldenvassdraget in the south-east of Norway. Currently it is impossible to distinguish river and brook lamprey genetically, so in this study they are treated as one species. I collected two independent water samples from 45 locations in Haldenvassdraget in an attempt to cover the whole river system. From the 45 locations 21 showed presence of lamprey, and 24 did not. Lamprey were primarily found in the upper parts of the river system and two rivers in the lower parts of the system, while no lamprey were found in the tributary Setten. In total, DNA-metabarcoding analyses on 16 samples failed, as such there is a possibility that the corresponding qPCR analyses on the same 16 samples also failed. Two generalised linear mixed model analyses were performed to test if the methods, qPCR and DNA-metabarcoding, differed in detection probability of lamprey. Sampling locations were added as a random factor. Water volume and temperature were added as fixed factors. The first analysis was done on the data set including results from both methods, excluding the 16 samples from DNA-metabarcoding that did not work as intended. The second analysis was done on the same data set where the 16 samples qPCR samples corresponding to the failed DNA-metabarcoding samples were removed. For the first analysis, detection probability did not differ between methods, but in the second analysis qPCR was shown to have a statistically significant higher detection probability than DNAmetabarcoding. Unexpectedly, for both analyses water volume had a negative impact on detection probability.

This study supports prior literature that qPCR in general is a more reliable method. I would also say that the study supports DNA-metabarcoding as a reliable alternative to qPCR when the objective is to investigate diversity and distribution for multiple freshwater fish. Also, in the case of lamprey, a species-specific protocol, where their life history and habitat preferences are taken into account, could be beneficial.

#### **Acknowledgments**

I had great hopes for my time as a master's student, but alas, with a pandemic and my health issues working against me, my grand plan did not come to fruition. Nevertheless, here we are, a master thesis delivered and one more student ready to join the grownups at work.

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As I now give thanks and show my appreciation, I am reminded of what dad always has told me. "Many will help you along the way, but never forget that it is still you who did the work". So, with that in mind I thank myself for never considering giving up. No matter what happens in the future I will always be proud of this accomplishment. In the end, I succeeded in what I set out to do.

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#### 1 Introduction

#### 1.1 Monitoring freshwater fish distribution

Sustaining biodiversity is becoming increasingly important to avert the Earth's trajectory towards the Sixth Mass Extinction (Cowie et al., 2022). Biodiversity is in danger of degradation at the genetic, species, and ecosystem level (Wrona et al., 2006), and 50% of all freshwater fish species are at threat from climate change worldwide (Ahmed et al., 2022). Monitoring must therefore take place to improve and ensure biodiversity. Environmental DNA (eDNA) monitoring is increasingly more popular as a tool to investigate biodiversity. Not only to find invasive (Fossøy et al., 2018) or endangered fish species (Thomsen et al., 2012), but also as a tool to investigate the distribution of local species (Rees et al., 2014; Rourke et al., 2022). eDNA offers a more objective method of monitoring as it does not rely on traditional taxonomic expertise in the field, long-term effort and does not disturb species being investigated (Fossøy et al., 2018; Strand et al., 2019). The method is based on that all organisms release DNA into their surroundings, and that it can be collected, for example by filtering water. DNA breaks down rapidly in nature. It is greatly influenced by temperature and water quality (Eichmiller et al., 2016), so a finding of one or more species indicates a high probability that they are present or have been in the area within a relatively short period of time (Balasingham et al., 2017). eDNA can therefore reflect the temporal and local diversity (Rees et al., 2014). The detection range of eDNA is disputed. In smaller rivers eDNA had a detection range of less than one km (Jane et al., 2015; Wilcox et al., 2016). In the outlet of lakes, dwelling species were detectable up to 9.1 km (Deiner & Altermatt, 2014). The further away from where the eDNA originated, the lower the probability of successful detection. eDNA can, however, potentially be detected at distances of up to 130 km (Pont, D. et al. 2018).

Compared to traditional monitoring methods eDNA methods are more sensitive and finds more species (Thomsen et al., 2012; Valentini et al., 2016), while also being less costly and time consuming. A study from 2018 investigating the downstream transportation ability of eDNA in rivers found that their sampling campaign of 12 days was equivalent to ten years of traditional sampling effort in order to reach an accurate image of fish biodiversity in the area (Pont et al., 2018).

In this study two eDNA methods were compared, qPCR and DNA-metabarcoding. Although there is comparable similarity between the methods (Yu et al., 2022), they are different in several ways. qPCR uses species-specific genetic markers to identify specific species, whereas DNA-metabarcoding uses a universal PCR primer pair to identify many species withing a taxonomical group, such as freshwater fish (Miya et al., 2020). qPCR as a monitoring method is more established than DNA-metabarcoding and perceived as more reliable (Bylemans et al., 2019; Lecaudey et al., 2019; Schenekar et al., 2020). Nevertheless, DNA-metabarcoding is seen as more suited for monitoring distribution of diversity in large river systems (Harper et al., 2018; Lecaudey et al., 2019).

As with all monitoring methods there are some drawbacks. eDNA methods are dependent on PCR where inhibiting substances in the water sample may affect the detection ability and potentially lead to false-negative results (Jane et al., 2015; Schrader et al., 2012). A known downside of DNA-metabarcoding is amplification bias (Bylemans et al., 2019; Elbrecht & Leese, 2015; Kelly et al., 2019; Tremblay et al., 2015). The MiFish primers are intended to cover a wide range of freshwater fish species (Miya et al., 2015). If there are species with relatively high abundance in the sample there is a chance that this abundance will block the detection of species with low DNA abundance. This amplification can also lead non-target-species DNA to be amplified (Alberdi et al., 2018; Gargan et al., 2022; Nichols et al., 2018; Piñol et al., 2019). For MiFish there might be difficulties with primer/template mismatches leading to underrepresentation of known species (Miya et al., 2020). Another downside to DNA-metabarcoding is the strong dependency on the comparison database; if a species DNA reference is missing then no match can occur. Moreover, there are also multiple families where it is not possible for DNA-metabarcoding to distinguish between individual species within the family (http://mitofish.aori.u-tokyo.ac.jp/species/all/).

#### 1.2 Little-studied fish in Norway: Lamprey

In this study, I will be investigating two species that are genetically indistinguishable (Docker, 2014; Zancolli et al., 2018) and are grouped as a paired species, river lamprey (*Lampetra fluviatilis*) and brook lamprey (*Lampetra planeri*). In this pair the river lamprey is the parasitic ancestor to the non-parasitic derivative brook lamprey (Docker, 2014; Zancolli et al., 2018). In Norway, river lamprey and brook lamprey are

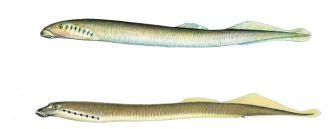


Figure 1 Top: river lamprey, bottom: brook lamprey. Owned by Jan Fekjan, license <u>CC BY-SA 4.0</u>

regarded as two different species, due to their clear morphological differences in their adult phase and differences in life cycle (Pethon, 2005; Spikkeland & Kasbo, 2014). Both lamprey species are known in Norway along with two other species which are not a part of this study, sea lamprey (*Petromyzon marinus*) and Arctic lamprey (*Lethenteron camtschaticum*). Lamprey distribution has not been sufficiently investigated in Norway (Hesthagen et al., 2021; Schartum & Kraabøl, 2013), and knowledge of lamprey distribution comes mainly from general fish diversity studies and bycatch data. In Haldenvassdraget, a river system located in eastern Norway, interest in lamprey species native to the region have begun to increase. A 2014 review about the distribution of river and brook lamprey in Haldenvassdraget reported that river lamprey was observed in, from north to south, Haretonelva (Ulviksjøen), Maltjennbekken, Langfossjøbekken (Longselva), Ørje, Fisma by Femsjøen (Rødselva). One brook lamprey was observed in Ørje, and unspecified lamprey species were observed in Midtskogvassdraget, Hemnessjøen, and Hafsteinselva (Spikkeland & Kasbo, 2014). A comprehensive review from 2021 continued to build on the knowledge of brook lamprey in Haldenvassdraget. Brook lamprey was observed in Bjørkelangen, Komnesbekken, Prestelva/Riselva, Gorobekken, Taraldrudelva (three observations), Rødenessjøen, Buerelva (three observations), Store Le, and Mellebyelva (Hesthagen et al., 2021). In addition to reviews three observations of river lamprey have been registered in Artsdatabanken (Artskart Artsdatabanken), which was not included in the reviews, were Mjerma, Femsjøen (Huitfeldt-Kaas, 1918), and Tista where it meets Iddefjorden. With their presence confirmed in Haldenvassdraget, it is reasonable to expect that I will be able to identify more locations where river and brook lamprey reside.

River and brook lamprey are suitable for this eDNA study because they represent species that are elusive and rare, resulting in them not being easily detected by traditional methods (Hesthagen et al., 2021; Huitfeldt-Kaas, 1918; Spikkeland & Kasbo, 2014). Based on their life cycle and habitat preferences it is probable that they appear throughout the river system even in places they have yet to be observed. Their life cycle and life history strategies are not well studied in Norway. Findings from a brook lamprey study from Telemark found that the life cycle and life history of brook lamprey in Norway follows the established knowledge of brook lamprey globally. I will therefore assume that any findings of brook lamprey globally will also hold true for brook lamprey in Haldenvassdraget, unless there is clear evidence to the contrary. There is a lot of research being done on lamprey on a world basis, and Docker (2014) has summarised and compiled the most comprehensive review of lamprey biology as of today, in the book Lampreys: Biology, Conservation and Control, Volume 1. From this book I have summarised relevant knowledge for this study. Both species begin their life cycle in freshwater, with a protracted larval phase, typically lasting between 3-7 years. During this phase they burrow down into soft sediment, which gives them protection from predators and allows them to filter feed on organic matter. When burrowed the larvae are difficult to observe and even more difficult to identify, leading to lampreys rarely being registered. Due to their worm-like shape, they are often mistaken as worms. They reside in streams and rivers, where the water is slow flowing and highly oxygenated. The larvae disperse largely based on factors such as changes in water velocity and water levels. However, they have some form of locomotion and actively seek better habitats for colonization. Important to note is that their ultimate upper incipient temperature is 30 °C, which is uncommonly hot in Haldenvassdraget. Lamprey larval habitat can be divided in to three types. Type I is the preferred habitat, located in the dispositional zone of rivers with the sediment being a mixture of sand and fine organic matter. Type II which has a much lower density of larval consists generally of shifting sands and potentially contains gravel. Type III, which is considered an unacceptable habitat for lamprey, consists of hard packed gravel, hardpan clay and bedrock. Compared to other lamprey species, larvae of river and brook lamprey are more commonly found in medium/coarse sand. After the larva phase metamorphosis occurs and the life cycle diverges between the parasitic river and non-parasitic brook lamprey. Once they have completed their metamorphosis, river lamprey enters their parasitic juvenile phase, and travel downstream to wider regions of large rivers where they feed until they have grown large and migrate

upstream to spawn as adults. Brook lamprey does not have a juvenile parasitic phase. Their sexual maturation commences during the metamorphosis from larvae to adult. As an adult they lose the ability to feed, and spawning happens soon after sexual maturity (Docker, 2014). Spawning behaviour begins in spring to early summer. Upstream migrations occur around 7-8°C for both species, and they both spawn at 10-11°C (Schartum & Kraabøl, 2013). Both species die about two weeks after spawning (Docker, 2014). Landlocked in Haldenvassdraget and unable to migrate to the ocean, river lamprey joins the brook lamprey as a potamodromous fish (Docker, 2014; Hesthagen et al., 2021). Other freshwater populations of river lamprey are also known from other countries (Degerman, 2009; Goodwin et al., 2006).

#### 1.3 Aim of study

The aim of this study is to assess detection probability of two different eDNA-based methods, the more established single-species qPCR and the newer, more complex multi-species DNA-metabarcoding, on elusive freshwater fish. River and brook lamprey in Haldenvassdraget were chosen as targeted species for the methods. In addition to assessing the methods, the results of this study will add to the knowledge of lamprey distribution within Haldenvassdraget.

#### 2 Materials and methods

#### 2.1 Study area

The study was conducted in the typical lowland river system, Haldenvassdraget (Figure 2). The system is a relatively large river-lake system in the south-east of Norway, with a length of 150 km, and a total catchment area of 1588 km2. Haldenvassdraget borders on three major river systems: in the north-east the Mangenvassdrag, in the west the large Glommavassdrag, and to the east Upperudsälven in Sweden (Johanson, 2002).



Figure 2. A: Map of northern Europe with Haldenvassdraget in blue. B: Map of Haldenvassdraget outlined in blue. Maps made in NVE Atlas

The main source of Haldenvassdraget springs from Dragsjøhaugen (268 m.a.s.l.) south of Årnes in Nes municipality. From Dragsjøhaugen the river system establishes the lakes Floangen and Floen, before it flows through Liermosen, one of Norway's largest peat bogs. It then continues through Bjørkelangen (124 m.a.s.l.), to Hølandselva ending in the lakes Skulerudsjøen and Rødenessjøen (118 m.a.s.l). The tributaries Hafsteinselva from Hemnessjøen (Øgderen) and Mjerma from Setten feed into Hølandselva. The geography of the river system changes after Skulerudsjøen, where it all the way to Halden can be categorized with short rivers connecting larger lakes. In quick succession there is Rødnessjøen (118 m.a.s.l.), Øymarksjøen (107 m.a.s.l), Aremarksjøen (105 m.a.s.l), Aspern (105 m.a.s.l) and Femsjøen (79 m.a.s.l). Haldenvassdraget ends in the river Tista which flows

out to Iddefjorden. The main river system is also joined by numerous smaller tributaries. (Johanson, 2002). Throughout the river system there are migration barriers for fish, notable ones include hydropower dams in Bjørkelangen, Ørje, Brekke, Strømsfoss and Tista (Johnsen & Vrålstad, 2009; Selbekk et al., 2020)

Due to human activity negatively affecting Haldenvassdraget continuously, measures to improve the water quality in Haldenvassdraget have been prioritised since the 1960s. The purpose of these measures put in place were to reduce nutrient supply to the river system from agriculture and treatment plants. In recent times monitoring has shown no clear trends in the development of water quality in the selected lakes. However, there is a clear connection between water quality and rainfall conditions, runoff, and the total phosphorus in the lakes (Greipsland et al., 2018).

The water quality is extremely varied in Haldenvassdraget. Water quality is quite poor in the upperparts, with the quality drastically improving downstream towards Iddefjorden. The upper parts of the main river and many of the tributaries are strongly affected by run-off from agriculture and households, causing a great deal of eutrophication, which gradually decreases through the system. The lakes filter out a lot of the dissolved excess nutrients. The characteristics of the lakes vary greatly, they come in all shapes and sizes, and nutrient richness (Spikkeland, 2014). Haldenvassdraget is located under the Marine limit, and heavily influenced by nutrient rich marine clay. Upstream from Ørje there is a steady supply of clay particles that are mainly responsible for the nutrient richness downstream. The amount of humus has increased in the whole river system the last 10 years. This is due to a decrease in acidification from acid rain (sulphur and nitrogen), and powerful precipitation resulting in leaching of humus into the river system. The variation in climate effects has a complex effect on the transport of total phosphorus and the algae development in the lakes, resulting in large variations in water quality from year to year (Greipsland et al., 2018).

Haldenvassdraget is considered species rich for Norway (Spikkeland, 2014). Fish migrating to Norway during the last ice age often ended up in Haldenvassdraget and did not migrate much further. The climate of the river system is also warmer than the rest of the country, which also contributes to a higher species richness (Huitfeldt-Kaas, 1918; Spikkeland, 2014). Fish species found in Haldenvassdraget include: sea lamprey, river lamprey, brook lamprey,

European eel (Anguilla anguilla), northern pike (Esox Lucius), Atlantic salmon (Salmo salar), sea trout (Salmo trutta), Arctic char (Salvelinus alpina), vendace (Coregonus Albula[FF1]), European whitefish (Coregonus lavaretus), European smelt (Osmerus eperlanus), roach (Rutilus rutilus), common dace (Leuciscus leuciscus), ide (Leuciscus idus), Eurasian minnow (Phoxinus phoxinus), rudd (Scardinius erythrophthalmus), asp (Aspius aspius), bleak (Alburnus alburnus), white bream (Blicca bjoerkna), freshwater bream (Abramis brama), crucian carp (Carassius carassius), common carp (Cyprinus carpio), burbot (Lota lota), threespined stickleback (Gasterosteus aculeatus), bullhead (Cottus gobio), alpine bullhead (Cottus poecilopus), European pearch (Perca fluviatilis), pike-perch (Stizostedion lucioperca), and ruffe (Gymnocephalus cernua) (Spikkeland 2014).

#### 2.2 eDNA sampling protocol

Haldenvassdragets' geography and local knowledge (from discussions with stakeholders) were considered when I was planning and conducting the study. Sampling locations were purposely selected to cover as much of the main river and tributaries as possible. Physical barriers, such as migration barriers hindering the movement of lampreys were also considered, in order to increase the possibility of detecting lamprey. Accessibility of the locations for sampling purposes were considered.

A total of 45 locations in Haldenvassdraget were visited and sampled from June 7<sup>th</sup> to June 13<sup>th</sup>, 2021 (Figure 2, Appendix table 1). The locations consisted of nine samples from the main river, and 36 samples from different tributaries. Samples were taken when good weather conditions had persisted for multiple days.

eDNA sampling was performed using sterile equipment at each location to hinder contamination. Water was collected from the water edge, to hinder contamination from footwear. To increase the likelihood of finding eDNA from lamprey residing above the sampling location, samples were acquired from turbulent or fast-moving water. Surgical gloves were always worn and disposed of between locations.

Two water samples were collected and filtrated at each site. The volume of water that was filtrated at different locations varied from 1 to 5 L due to the particles filling up the filter at

different rates at each location. A 1000 mL water sampler (plastic bucket) was used to collect running water, a deposable silicone tube was used to pump the water through a NatureMetrics capsule filter, consisting of two filters, a 5.0 µm glass fibre filter, and a 0.8 µm PES main filter, by the help of a peristaltic pump (Bürkle Vampire). To avoid damage to the filter, the speed of the pump was set to 1. At the end of each filtration, the capsule was emptied of water and dried.

A syringe with preservative buffer (ATL-buffer, Qiagen) was screwed onto the inlet of the capsule, holding the syringe and capsule upright so the outlet pointed up, the buffer was slowly introduced to the filter, until the filter was completely saturated with buffer, indicated with a droplet emerging on the outlet of the filter. About 1.5 mL ATL-buffer (Qiagen) was added to each capsule. A luer cap was then screwed on the inlet while the filter was pointing up. After that, with the filter pointing down, the syringe was unscrewed and a luer cap was screwed on the outlet. The filters were then labelled with number, date, location and station in chronological order.

Temperature was measured using a digital thermometer at each location after the samples were collected.

The filters were stored at dark conditions and room temperature (above 10°C), to prevent crystallization, due to the preservation buffer in the syringe being temperature sensitive. The samples were then transported to the Centre for Biodiversity Genetics (NINAGEN) at NINA in Trondheim for genetic analyses.

#### 2.3 DNA – extraction

DNA-extraction, qPCR and library prep for DNA-metabarcoding were done by professional lab technicians at the Norwegian Centre for Biodiversity Genetics (NINAGEN) in Trondheim. For the DNA-extraction, 450  $\mu$ L Proteinase-K (Qiagen) was added to the sampling tubes and incubated overnight at 56°C. DNA was isolated using NucleoSpin Plant II Midi kit (Macherey-Nagel), following the manufacturers protocol except that Qiagen buffers were used instead of those supplied with the kit. Extracted DNA was eluted in 200  $\mu$ L AE buffer. Samples were re-eluted for maximizing the output of DNA.

#### 2.4 qPCR analysis

The species-specific genetic markers cannot differentiate between river and brook lamprey as of now, they are therefore considered ecotypes when analysed using quantitative PCR (qPCR) for detection of the target species using the LampATPase6\_1 assay (Zancolli et al., 2018). Each qPCR-reaction had a total volume of 30 µL which included 15 µL TaqMan Fast Advanced Master Mix (ThermoFisher Scientific), 0.9 µM of forward, reverse primer and probe, 4.5 dH20 and 5 µL DNA-template. PCR-conditions started with an onset of 50°C for 2 min and 95°C for 10 min, followed by 45 cycles of 95°C for 15 sec and 60°C for 90 sec, and were finalized with 72°C for 10 min. All samples were analysed using a QuantStudio 5 qPCR-machine (ThermoFisher Scientific), and all samples were analysed in triplicates where only samples showing at least 2 out of 3 positive results were considered positive for the target species.

#### 2.5 Metabarcoding analysis

For DNA-metabarcoding of fish, the 12S region was amplified using the MiFish-U-F and MiFish-U-R primers (Miya et al., 2015). PCR was conducted in 25 μL volumes containing: 1X KAPA HiFi HotStart Ready Mix, 0.3 μM of each primer, and 2.5 μL of 10 ng/μL template DNA. The PCR conditions were as follows; first a denaturation step of 95°C for 3 minutes, followed by 35 cycles of: 98°C for 20 seconds, annealing at 65°C for 15 seconds, and elongation at 72°C for 15 seconds. The PCR were finalized with 72°C for 5 minutes.

Amplicons were quality checked on a Tape Station (Agilent 4200) and cleaned with magnetic beads (MAG-BIND RXN PURE PLUS) to remove fragments <200bp after each PCR. Indexes for Illumina DNA/RNA UD indices were added to the 5' and 3' ends of the amplicons according to the manufacturer's instructions. Amplicons were diluted to 6 ng/μL, and magnetic beads were used to remove fragments <500bp. Amplicons were pooled in equimolar amounts and sequenced on an Illumina NovaSeq 6000 platfrom at the Norwegian Sequencing Centre (NSC) in Oslo.

#### 2.6 DNA – metabarcoding bioinformatics

Demultiplexing was performed on the Illumina NovaSeq platform. Primer removal was conducted with cutadapt v. 1.9.1 requiring a minimum length match of 17 bp with 0.15 expected errors. Quality filtering, error correction, merging, mapping and chimera removal were conducted in DADA2 (Callahan et al., 2016) to generate ASV (Amplicon Sequence Variants). ASV, the quality assured DNA-sequences, were used to separate the species, as well as investigation genetic variation within species. Taxonomy was assigned using BLASTN comparisons to the GenBank database. A minimum of 97% identity and 90% coverage with a reference sequence were required for a successful assignment at the species level.

#### 2.7 Statistical analyses

I used generalised linear mixed models (GLMM) for testing whether the two methods, qPCR and DNA-metabarcoding, differed in detection probability of lamprey. I included water volume and temperature in addition to method as fixed factors. Sampling location was added as a random factor to control for non-independence of samples collected at the same site. Water volume was considered due to larger water volumes previously being found to increase detection probability (Agersnap et al., 2017; Muha et al., 2019). Water temperature has been shown to affect the amount of eDNA released by fish, where more eDNA was released at higher temperatures (Lacoursière-Roussel et al., 2016; Wacker et al., 2019). I used the package lmerTest (Kuznetsova et al., 2017) in the statistical software R (Team, 2021) to run the mixed models.

#### 3 Results

#### 3.1 qPCR

All samples were analysed in triplicates with a threshold set of at least 2 out of 3 positive results to be considered positive for lamprey. As shown in Table 1, 37 of 90 samples showed presence of lamprey, resulting in lamprey being detected at 20 of 45 location. 52 of 90 samples showed no presence of lamprey, resulting in lamprey not being detected at 25 of 45 location. Sample 14B Upper Mjerma failed, due to the lid of the tube that the sample was in, opened while in the centrifuge, and was lost.

At some locations there were differences between the two replicate samples. Hemneselva (07) showed some differences in presence, with sample A showing presence in 3/3 replicates and sample B showing presence in 2/3 replicates. There was also a difference within Lower Mjerma (15), where sample A indicated no presence (1/3 positive replicates) and sample B showing presence (3/3 positive replicates). Other locations where there were differences between samples was Dalselva and Lower Hafsteinselva, where sample A showed no presence (0/3 positive replicates) and samples B showed presence (3/3 positive replicates). If either one of the two samples are positive, I accept the location as positive for presence of lamprey.

For qPCR all positive controls were positive and negative controls were negative.

Table 1 Results from qPCR analysis showing sample ID, locality, number of replicates and how many of the replicas tested positive for lamprey DNA.  $C_T$  is PCR cycles. Lower  $C_T$  Mean equates to higher concentration of DNA.  $C_T$  SD is the standard deviation of  $C_T$ .

Sample ID	Locality	No. replicates	No. positive replicates	Ct Mean	Ct SD
01A	Ulviksjøen	3	1	37.10	
01B	Ulviksjøen	3	0		
02A	Eidsbekken	3	3	27.99	0.13
02B	Eidsbekken	3	3	27.96	0.05
03A	Lierelva	3	3	39.34	1.42
03B	Lierelva	3	3	33.93	0.89
04A	Snartjern	3	3	27.93	0.13
04B	Snartjern	3	3	28.18	0.08

05A	Malttjernbekken	3	3	28.97	0.03
05B	Malttjernbekken	3	3	29.41	0.13
06A	Dalselva	3	0		
06B	Dalselva	3	3	27.80	0.24
07A	Hemneselva	3	3	34.46	0.93
07B	Hemneselva	3	2	37.59	0.77
08A	Lower Hafsteinselva	3	0		
08B	Lower Hafsteinselva	3	3	28.25	0.14
09A	Hølandselva	3	3	29.67	0.19
09B	Hølandselva	3	3	31.18	0.15
10A	Upper Hafsteinselva	3	3	24.64	0.06
10B	Upper Hafsteinselva	3	3	31.08	8.64
11A	Setta	3	0	31.00	0.01
11B	Setta	3	0		
12A	Langebruslora	3	0		
12B	Langebruslora	3	0		
13A	Langtjen	3	0		
13A	Langtjen	3	1	37.69	
14A	Upper Mjerma	3	0	37.09	
14A 14B					
14B	Upper Mierma	3	1	24.00	
	Lower Mjerma			34.89	0.25
15B	Lower Mjerma	3	3	28.42	0.25
16A	Nautebrofoss	3	3	28.49	0.32
16B	Nautebrofoss	3	3	26.87	0.26
17A	Østenbyelva	3	0		
17B	Østenbyelva	3	0		
18A	Taraldruelva	3	3	30.97	4.36
18B	Taraldruelva	3	3	25.40	0.14
19A	Åsebyelva	3	3	35.62	0.98
19B	Åsebyelva	3	3	35.22	0.67
20A	Langnes	3	3	28.62	0.05
20B	Langnes	3	3	29.04	0.18
21A	Engerelva	3	3	7.626	0.07
21B	Engerelva	3	3	28.08	0.10
22A	Ørje	3	0		
22B	Ørje	3	1	44.14	
23A	Bøenselva	3	0		
23B	Bøenselva	3	0		
24A	Gunnengbekken	3	3	37.60	1.30
24B	Gunnengbekken	3	3	36.33	1.45
25A	Skinnarbutjenn	3	0		
25B	Skinnarbutjenn	3	0		
26A	Langetjernelva	3	0		
26B	Langetjernelva	3	0		
27A	Fangebekken	3	3	32.34	0.25
27B	Fangebekken	3	3	31.71	0.08

28A	Fossby	3	1	28.65	
28B	Fossby	3	0		
29A	Skolleborg	3	0		
29B	Skolleborg	3	0		
30A	Verksbrua	3	0		
30B	Verksbrua	3	0		
31A	Tenebekken	3	0		
31B	Tenebekken	3	0		
32A	Strømsfoss	3	0		
32B	Strømsfoss	3	0		
33A	Lielva	3	3	34.92	1.19
33B	Lielva	3	3	34.11	0.22
34A	Valbyelva	3	0		
34B	Valbyelva	3	0		
35A	Holmegilelva	3	0		
35B	Holmegilelva	3	0		
36A	Kverntjern	3	0		
36B	Kverntjern	3	0		
37A	Stenselva	3	0		
37B	Stenselva	3	0		
38A	Ganerødelva	3	0		
38B	Ganerødelva	3	0		
39A	Mellebyelva	3	0		
39B	Mellebyelva	3	0		
40A	Rødselva	3	0		
40B	Rødselva	3	1	41.71	
41A	Rjørelva	3	0		
41B	Rjørelva	3	0		
42A	Asakbekken	3	3	27.94	0.09
42B	Asakbekken	3	3	27.98	0.12
43A	Tistedalsfoss	3	3	37.23	1.57
43B	Tistedalsfoss	3	3	40.99	2.63
44A	Skåningsfoss	3	0		
44B	Skåningsfoss	3	0		
45A	Porsnes sluser	3	0		
45B	Porsnes sluser	3	0		

#### 3.2 DNA – metabarcoding

There was large variation in the number of total reads for DNA-metabarcoding between samples (Appendix table 1). Percentage of detected lamprey relative reads varied between less than 1% to 47% (Table 2). Most samples were acceptable for the number of reads, but for the following location the total reads were quite low seen in context to lamprey relative reads, like in Asakbekken (42B) 0.47/5480.

As shown in Table 2, 31 of 90 samples were positive for lamprey, indicating that lamprey was present in 19 of 45 locations. Lamprey was not present in 43 of 90 samples, indicating that lamprey was not present in 24 of 45 locations. 16 of 90 samples failed, but there were only two locations where both sample A and B failed. This equates to 17. 7 % of the samples not working correctly. These samples showed presences for bacteria and fungi but lacked freshwater fish DNA.

Within 18 locations there were differences between the two replicate samples. Both replicate samples failed in Upper Mjerma (14) and Holmegilelva (35). For Rødselva (40), one sample failed and the other showed presence of lamprey. In 11 locations one sample failed, and the other showed no presence of lamprey. The locations were Setta (11), Østenbyelva (17), Langtjernelva (26), Fossby (28), Skolleborg (29), Verksbrua (30), Tenebekken (31), Kverntjern (36), Stenselva (37), Mellebyelva (39) and Skåningsfoss (44).

Both positive and negative controls for river/brook lamprey worked.

Table 2 Number of reads per sample after DNA-metabarcoding and bioinformatic analyses (see more in Appendix table 1). No. fish reads, is the total number of fish reads in the sample. No. lamprey reads, is how many reads of the No. fish reads were lamprey reads. Lamprey relative reads is relative lamprey reads. And result show how many samples tested positive for lamprey.

Sample ID	Locality	No. fish reads	No. lamprey reads	Lamprey relative reads	Result
01A	Ulviksjøen	123143	0	0	NEG
01B	Ulviksjøen	108306	0	0	NEG
02A	Eidsbekken	524914	3362	0.01	POS
02B	Eidsbekken	554446	4463	0.01	POS
03A	Lierelva	128354	4991	0.04	POS
03B	Lierelva	136224	1522	0.01	POS
04A	Snartjern	296624	2855	0.01	POS

04B	Snartjern	339131	1969	0.01	POS
05A	Maltjennbekken	137131	2717	0.02	POS
05B	Maltjennbekken	173915	1614	0.01	POS
06A	Dalselva	117960	11297	0.10	POS
06B	Dalselva	58284	2891	0.05	POS
07A	Hemneselva	141659	95	0.00	POS
07B	Hemneselva	36005	0	0	NEG
08A	Lower Hafsteinselva	18588	0	0	NEG
08B	Lower Hafsteinselva	20196	654	0.03	POS
09A	Hølandselva	61200	1094	0.02	POS
09B	Hølandselva	32323	54	0.00	POS
10A	Upper Hafsteinselva	252563	33295	0.13	POS
10B	Upper Hafsteinselva	149066	21099	0.14	POS
11A	Setta	84863	0	0	NEG
11B	Setta	0	NA	NA	NA
12A	Langebruslora	30620	0	0	NEG
12B	Langebruslora	26481	0	0	NEG
13A	Langtjen	18980	0	0	NEG
13B	Langtjen	44228	0	0	NEG
14A	Upper Mjerma	0	NA	NA	NA
14B	Upper Mjerma	0	NA	NA	NA
15A	Lower Mjerma	143222	672	0.00	POS
15B	Lowe Mjerma	96479	2145	0.02	POS
16A	Nautebrofoss	59424	4514	0.08	POS
16B	Nautebrofoss	93011	5351	0.06	POS
17A	Østenbyelva	0	NA	NA	NA
17B	Østenbyelva	54421	0	0	NEG
18A	Taraldruelva	81966	16275	0.20	POS
18B	Taraldruelva	78025	21135	0.27	POS
19A	Åsebyelva	14715	139	0.01	POS
19B	Åsebyelva	7642	0	0	NEG
20A	Langnes	126497	429	0.00	POS
20B	Langnes	168234	0	0	NEG
21A	Engerelva	163123	4613	0.03	POS
21B	Engerelva	195665	3117	0.02	POS
22A	Ørje	38584	0	0	NEG
22B	Ørje	33260	0	0	NEG
23A	Bøenselva	46218	0	0	NEG
23B	Bøenselva	36906	0	0	NEG
24A	Gunnengbekken	11597	24	0.00	POS
24B	Gunnengbekken	20850	0	0	NEG
25A	Skinnarbutjenn	52248	0	0	NEG
25B	Skinnarbutjenn	82281	0	0	NEG
26A	Langetjernelva	0	NA	NA	NA
26B	Langetjernelva	91083	0	0	NEG
27A	Fangebekken	195522	0	0	NEG
	9				

27B	Fangebekken	145248		18		0.00	POS
28A	Fossby	206737		0		0	NEG
28B	Fossby	0	NA		NA		NA
29A	Skolleborg	543		0		0	NEG
29B	Skolleborg	8255		0		0	NEG
30A	Verksbrua	0	NA		NA		NA
30B	Verksbrua	61726		0		0	NEG
31A	Tenebekken	0	NA		NA		NA
31B	Tenebekken	3537		0		0	NEG
32A	Strømsfoss	75699		0		0	NEG
32B	Strømsfoss	40798		0		0	NEG
33A	Lielva	160183		0		0	NEG
33B	Lielva	221201		0		0	NEG
34A	Valbyelva	155144		0		0	NEG
34B	Valbyelva	68240		0		0	NEG
35A	Holmegilelva	0	NA		NA		NA
35B	Holmegilelva	0	NA		NA		NA
36A	Kverntjern	40398		0		0	NEG
36B	Kverntjern	0	NA		NA		NA
37A	Stenselva	20758		0		0	NEG
37B	Stenselva	606		0		0	NEG
38A	Ganerødelva	12855		0		0	NEG
38B	Ganerødelva	24760		0		0	NEG
39A	Mellebyelva	41782		0		0	NEG
39B	Mellebyelva	0	NA		NA		NA
40A	Rødselva	0	NA		NA		NA
40B	Rødselva	48345		415		0.01	POS
41A	Rjørelva	39359		0		0	NEG
41B	Rjørelva	14532		0		0	NEG
42A	Asakbekken	97710		3879		0.04	POS
42B	Asakbekken	5480		2594		0.47	POS
43A	Tistedalsfoss	19947		0		0	NEG
43B	Tistedalsfoss	19686		0		0	NEG
44A	Skåningsfoss	0	NA		NA		NA
44B	Skåningsfoss	7980		0		0	NEG
45A	Porsnes sluser	31134		0		0	NEG
45B	Porsnes sluser	19765		0		0	NEG

# 3.3 Comparing qPCR and metabarcoding for detecting lampreys

The combined the results of qPCR (Table 1) and DNA-metabarcoding (*Table 2*) results in 21 locations of lamprey presence and 24 locations of lamprey absence (Table 3). If at least one sample per location tested positive for lamprey in either of these methods, lamprey was considered present in that location.

DNA-metabarcoding showed no fish in 16 samples, but OD measurements, a measurement used to detect contaminants, show decent quality DNA in the DNA-extracts (Appendix table 2). It is unlikely that there were no fish in these locations, and the results from the qPCR analyses also came back negative for all 16 samples. It is therefore not possible to differentiate whether these samples also did not work for the qPCR analyses, or if they were just negative for the presence of lamprey.

Table 3 Summery of Table 1 and 2 of lamprey in Haldenvassdraget based on eDNA-samples analysed using qPCR (sample A and B) or DNA-metabarcoding – Bar (sample A and B). If lamprey was detected using one of the methods, presences was accepted as true for the location.

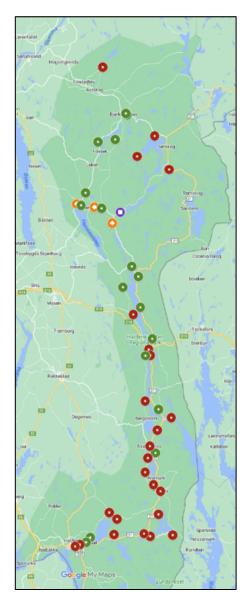
#	Locality	qPCR A	qPCR B	Bar A	Bar B	Presence
1	Ulviksjøen	NEG	NEG	NEG	NEG	ABSENT
2	Eidsbekken	POS	POS	POS	POS	PRESENT
3	Lierelva	POS	POS	POS	POS	PRESENT
4	Snartjern	POS	POS	POS	POS	PRESENT
5	Maltjennbekken	POS	POS	POS	POS	PRESENT
6	Dalselva	NEG	POS	POS	POS	PRESENT
7	Hemneselva	POS	POS	POS	NEG	PRESENT
8	Lower Hafsteinselva	NEG	POS	NEG	POS	PRESENT
9	Hølandselva	POS	POS	POS	POS	PRESENT
10	Upper Hafsteinselva	POS	POS	POS	POS	PRESENT
11	Setta	NEG	NEG	NEG	-	ABSENT
12	Langebruslora	NEG	NEG	NEG	NEG	ABSENT
13	Langtjenn	NEG	NEG	NEG	NEG	ABSENT
14	Upper Mjerma	NEG	-	-	-	ABSENT
15	Lower Mjerma	NEG	POS	POS	POS	PRESENT
16	Nautebrofoss	POS	POS	POS	POS	PRESENT
17	Østenbyelva	NEG	NEG	-	NEG	ABSENT

18	Taraldrudelva	POS	POS	POS	POS	PRESENT
19	Åsebyelva	POS	POS	POS	NEG	PRESENT
20	Langnes	POS	POS	POS	NEG	PRESENT
21	Engerelva	POS	POS	POS	POS	PRESENT
22	Ørje	NEG	NEG	NEG	NEG	ABSENT
23	Bøenselva	NEG	NEG	NEG	NEG	ABSENT
24	Gunnengbekken	POS	POS	POS	NEG	PRESENT
25	Skinnarbutjern	NEG	NEG	NEG	NEG	ABSENT
26	Langetjernelva	NEG	NEG	-	NEG	ABSENT
27	Fangebekken	POS	POS	NEG	POS	PRESENT
28	Fossby	NEG	NEG	NEG	-	ABSENT
29	Skolleborg	NEG	NEG	-	NEG	ABSENT
30	Verksbrua	NEG	NEG	-	NEG	ABSENT
31	Tenebekken	NEG	NEG	-	NEG	ABSENT
32	Strømsfoss	NEG	NEG	NEG	NEG	ABSENT
33	Lielva	POS	POS	NEG	NEG	PRESENT
34	Valbyelva	NEG	NEG	NEG	NEG	ABSENT
35	Holmegilelva	NEG	NEG	-	-	ABSENT
36	Kverntjern	NEG	NEG	NEG	-	ABSENT
37	Stenselva	NEG	NEG	NEG	-	ABSENT
38	Ganerødelva	NEG	NEG	NEG	NEG	ABSENT
39	Mellebyelva	NEG	NEG	NEG	-	ABSENT
40	Rødselva	NEG	NEG	-	POS	PRESENT
41	Rjørelva	NEG	NEG	NEG	NEG	ABSENT
42	Asakbekken	POS	POS	POS	POS	PRESENT
43	Tistedalsfoss	POS	POS	NEG	NEG	PRESENT
44	Skåningsfoss	NEG	NEG	-	NEG	ABSENT
45	Porsnes sluser	NEG	NEG	NEG	NEG	ABSENT

Considering Table 3 and Figure 3 it is clear that there is no presence of lamprey eDNA at the sampling starting point Ulviksjøen, but it was detected in the main river leading to Bjørkelangen, and the tributaries connecting to the lake. The main rivers connecting to Hemnessjøen and Rødenessjøen through Hølandselva also showed presence. One tributary, west connecting to Rødenessjøen showed no presence. No lamprey was detected in any of the location in the tributary Setten but lamprey was detected below the waterfall from Setten connecting to the main river Hølandselva. No lamprey was detected in Ørje river, connecting Rødenessjøen and Øymarksjøen. Two tributaries connecting to Øymarksjøen show lamprey presence, one on the northwest side of the lake and the other on the east side. Only one tributary connecting to Aremarksjøen, in the northeast, showed lamprey presence, no presence

was detected in the main river. No presence was detected throughout Aspen, Stenselva and tributary Ganerødelva, connecting to Store Erte. In Femsjøen, only the tributary Asakbekken showed lamprey presence. The last location where lamprey was present was Tistedalsfoss. No presence was found downstream after Porsnes sluser, which marks the end of my sampling of Haldenvassdraget.

Based on the map (Figure 3 C) it would seem that the upper parts of Haldenvassdraget contain more lamprey than in the lower part of the system, and that there is no lamprey in the northeast tributary Setten.



A: qPCR result



B: DNA-metabarcoding result

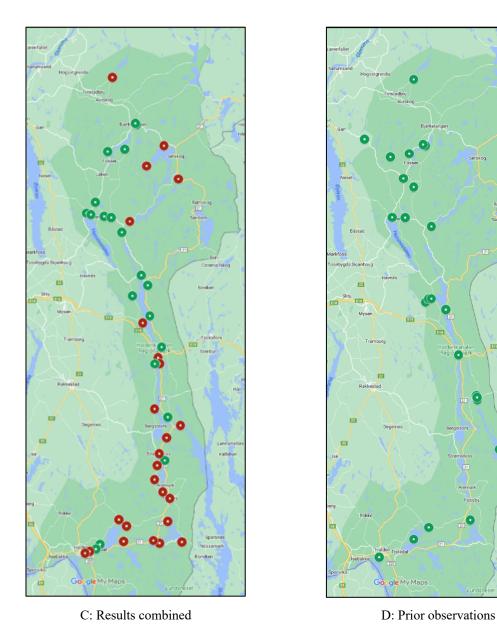


Figure 3 The maps show the distribution of lamprey presence in Haldenvassdraget for A: qPCR, B: DNA-metabarcoding and C: qPCR and DNA-metabarcoding combined. Green: lamprey DNA present in both samples, Red: lamprey DNA absent from both samples, Orange: lamprey DNA present in one sample and absent in the other, Purple: absence of lamprey DNA in one sample and one failed sample, Blue: presence of lamprey DNA in one sample and one failed sample, Black: both samples failed. In C: Green: lamprey DNA present, and Red: lamprey DNA absent. D: Prior observation from (Hesthagen et al., 2021; Spikkeland & Kasbo, 2014) and Artsdatabanken (Artskart Artsdatabanken). Maps made in Google.

#### 3.4 Probability of detecting lamprey

To test if there was a difference in detection probability of lamprey between the two methods, qPCR and DNA-metabarcoding, generalized linear mixed models (GLMM) was used. Analysis was run on two different sets of data, henceforth data set A and data set B. In A, the 16 DNA-metabarcoding samples which had failed were removed and in B, in addition to the removal of failed DNA-metabarcoding samples, the corresponding qPCR results, which were all negative, were removed.

The differences in detection probability of lamprey between the different methods had no significant differences in A (Table 4A), however in B qPCR was statistically significantly better than DNA-metabarcoding (Table 4B). Both models show that water volume had a negative effect on detection probability, and water temperature had no significant effect.

Table 4 Model output from GLMM on the probability of detecting lamprey. The eDNA detection methods (qPCR and DNA-metabarcoding) are the main factors, with water temperature and volume as co-variables. A is testing with omitted failed DNA-metabarcoding results, and B is testing with omitted corresponding qPCR results.

Α	Estimate	SE	Z	Р
(Intercept)	8.72	7.71	1.13	0.26
DNA-method (qPCR)	0.88	0.63	1.39	0.17
Water temperature	-0.42	0.44	-0.96	0.34
Water volume	-1.9	0.79	-2.42	0.016
В	Estimate	SE	Z	Р
B (Intercept)	Estimate 2.71	<b>SE</b> 3.24	<b>Z</b> 0.84	<b>P</b> 0.40
_			_	-
(Intercept)	2.71	3.24	0.84	0.40

#### 4 Discussion

# 4.1 Detection performance of qPCR and DNA-metabarcoding

In this study I compared the detection sensitivity of qPCR and DNA-metabarcoding for lamprey in 45 locations, with two samples collected from each location. I compared if the samples were positive or negative for lamprey DNA within each method, and then compared the result from both methods to investigate presence and absence of lamprey for each location. The results suggest that both methods can detect lamprey, and the results from both methods show a similar spatial distribution for lampreys (Table 3, Figure 3).

To compare the methods, I checked for statistically significant differences by running a GLMM on the probability of detecting lamprey, as well as covariables, water temperature and water volume (Table 4). In DNA-metabarcoding, 16 of 90 samples failed, having no fish matches, suggesting no freshwater fish DNA in the samples. There was however DNA-sequences from non-target species, mainly represented by fungi DNA and bacteria DNA. There is no way of knowing what caused the samples to fail during analysis. It could be anything from amplification bias (Bylemans et al., 2019; Elbrecht & Leese, 2015; Kelly et al., 2019; Tremblay et al., 2015), amplification of non-target-species DNA (Alberdi et al., 2018; Gargan et al., 2022; Nichols et al., 2018; Piñol et al., 2019), mismatches with primer/template (Miya et al., 2020) or even PCR inhibition (Jane et al., 2015). These 16 DNA-metabarcoding results were discarded and the remaining 74 along with the 90 qPCR results comprised dataset A. The differences in detection probability of lamprey between the different methods had no significant differences in dataset A (Table 4A).

Interestingly, for all 16 samples of DNA-metabarcoding that failed to find fish DNA, the qPCR of those samples were negative for lamprey. This means that the qPCR for these samples might also have failed. There is no way of telling whether these qPCR samples are in fact negative, or if there were problems with the analyses. This uncertainty changes the main conclusion of my work. Removing these 16 samples also from qPCR results, the remaining 74 results along with the 74 DNA-metabarcoding results comprised dataset B. In this dataset there was a significant difference in detection probability of lamprey between the different methods, favouring qPCR over DNA-metabarcoding (Table 4B). This finding aligns with the

consensus that qPCR is a more sensitive method (Bylemans et al., 2019; Lecaudey et al., 2019; Schenekar et al., 2020). For a more reliable comparison, a positive control for qPCR analysis for a targeted species on each sample would be beneficial.

Both models showed that water volume had a negative effect on detection probability, and water temperature had no significant effect. This result is initially counter intuitive, due to higher water volume leading to more total eDNA in samples for a given concentration of eDNA (Bessey et al., 2020). There is no way of knowing why water volume had a negative effect on detection probability of lamprey in Haldenvassdraget. The ability of eDNA to attach to or be contained within particles is known to affect the chance to detect targeted species (Turner et al., 2014). An unknown factor is how well lamprey DNA attaches to organic materials, materials which might clog filters faster, hence leading to less water being filtered. From my combined results of locations of lamprey presence, the lampreys in Haldenvassdraget resided in areas with higher eutrophication and humus, so the preferred habitat by lampreys in Haldenvassdraget might contain more organic matter that lamprey DNA can attach to and clog filters faster.

A more general reason for why large water volume is not strictly needed, is that eDNA analyses methods are quite sensitive, so reaching the minimum required amount of DNA for eDNA analyses methods to conclude a positive result, is well within reason for 1L sample volume as it is for 5L sample volume. Previous studies have cited both 1L and 5L as sufficient volumes when sampling for eDNA (Harper et al., 2018; Yu et al., 2022). The volumes of my samples, 1L to 5L, might also not be a large enough sampling range to conclude anything decisively and the sample volumes are not distributed equally over the range, as there are much more sample volumes of 1L, or close to, than there are of sample volumes of 5L. My dataset of sample volumes being heavily skewed to the lower end, not only makes results more biased, but also means that upscaling or down scaling sample volumes, or changing sample location, will not necessarily yield similar connection of lower volume increasing detection chance.

Excluding the failed samples and looking at the remaining 74 samples, 37 out of 74 qPCR results were positive for lamprey, resulting in lamprey being detected at 20 locations (Table 1). Furthermore, 31 out of 74 DNA-metabarcoding results were positive for lamprey, resulting in lamprey being detected at 19 of 45 locations (Table 2).

Three locations stood out where qPCR and DNA-metabarcoding showed contradictive results. These locations were Lielva (33), Rødselva (40) and Tistedalsfoss (43). For Lielva (33), and Tistedalsfoss (43) qPCR detected lamprey while DNA-metabarcoding did not. For Rødselva (40) qPCR detected no lamprey DNA, however DNA-metabarcoding did detect lamprey. This is unexpected as qPCR is regarded as being more sensitive. Previous studies support qPCR having a higher detection sensitivity than DNA-metabarcoding when using fish as targeted species (Bylemans et al., 2019; Lecaudey et al., 2019; Schenekar et al., 2020). This is also true for other comparison studies where aquatic animals, such as great crested newt (*Triturus cristatus*) and Mediterranean fanworm (*Sabella spallanzanii*), were used (Harper et al., 2018; Wood et al., 2019).

With some samples failing it is important to assess the quality of this study. The sampling protocol I used while sampling was developed by NINA with the intent to standardise sampling, but also to make it easy enough to be used by citizen scientists without formal training. There was however there is no negative control.

The possibility of false positive and false negative results was considered, and steps were taken to minimise the risk of them occurring. For both qPCR and DNA-metabarcoding, water was used as a negative control and lamprey DNA was used as a positive control. The positive and negative controls worked. Despite this, it seems that some of the negative controls for the DNA-metabarcoding were slightly contaminated, but it had no effect on the final result. To limit uncertainty for qPCR results, a criterion was set defining lamprey presence if 2 out of 3 results showed presence of lamprey DNA. These criteria make it less probable that the results are false positive or false negative. However, thresholds affect detection (Harper et al., 2018), so my study might have had a too low or too high of a threshold.

I sampled in June, when lamprey are known to be active in Norway (Schartum & Kraabøl, 2013), so eDNA concentration should have be relatively high compared to other seasons. However, it might have been good to sample in a season with low water as this is shown to increase detection of eDNA (Bylemans et al., 2019).

A central part of this study was to compare the two eDNA methods, qPCR and DNA-metabarcoding, and an essential part of that comparison is the cost. The collection kits and lab analyses for qPCR was NOK 1500 per kit/filter. Thus for 2 filters per location, it was NOK 3000 per location. The cost applied to one genetic marker (one species). An additional cost of

NOK 700 could be added per species per filter if it was wanted to test for several species (such as if test for sea trout in addition to river trout). DNA – metabarcoding was a little bit more expensive. It was NOK 3500 per kit/filter, including collection kits (field equipment). Thus for 2 filters per location, it was NOK 7000 per location with 2 filters.

An added benefit to DNA-metabarcoding, is that it can investigate fish species diversity within a single sample with MiFish (Bessey et al., 2020; Ruppert et al., 2019). Although it is not a focus in my study, multiple species were found at each location (see Appendix table 3). My study had no need for the additional findings of other fish species, but the findings are beneficial for further research and management of Haldenvassdraget. An example of a species I found, in addition to, is eel, which is a species of interest in Haldenvassdraget.

Some of the biggest hinders in capturing optimal eDNA samples were the widths of the rivers, water volume and the lack of points with high turbulence that would have mixed the water well enough to get a good representation of the biodiversity. The uncertainty around how far eDNA can reliably be detected is also a hindrance. That being said it would seem that lamprey eDNA do not travel extreme distances (130 km) in Haldenvassdraget. If that was the case, I would have detected much more lamprey eDNA in the main river throughout the system.

# 4.2 Added knowledge to the distribution of lamprey in Haldenvassdraget

In addition to comparing the two eDNA methods, my study has added to the knowledge of distribution of river and brook lamprey in Haldenvassdraget. A sample absent of the targeted species eDNA, in this case lamprey eDNA, is not a guaranteed lack of presence of the targeted species at sampling site. Just because I did not find lamprey at a location, does not mean they were not there. This is supported by cross-referring my result with prior observations. Lamprey absence in samples does not prove actual absence (Roussel et al., 2015).

Out of the 45 locations I sampled from, 10 of them overlapped with previously reported observations of lamprey (Figure 3 C and D). Some of the locations do not have enough detail about exact placement making it difficult to find the exact spot the observation was made, so

some of the observation points on the map are estimates based on my knowledge of Haldenvassdraget. I was able to confirm presence of lamprey in Maltjennbekken (5), Hemneselva (7), Lowe (8), Upper Hafsteinselva (10) and Rødselva (40). There was also an observation from Mjerma where I sampled, I assumed it was around Lower Mjerma (15), due to Lundefossen separating Upper Mjerma (14) and Setten from the lower parts of Mjerma. I was however not able to confirm the presence of lamprey in Ulviksjøen (1), Ørje (22), and Mellebyelva (39) even though lamprey has been observed at some point. The observation Ulviksjøen (1) is from 1995 so it might be possible that the lamprey population is no longer there, if that is the case they probably moved, which lamprey are known to do (Dekker). However, it is also probable that I just did not capture any lamprey eDNA when I was sampling. For Ørje (22), presence has been confirmed for both river and brook lamprey from 2010 (2014), so it is most likely that I did not capture any eDNA. Considering the width of the river it was not feasible to acquire a well-mixed water sample. Presence of lamprey in Mellebyelva (39) was confirmed in 2015, and, again, the lamprey populations might have moved, or the more likely scenario, I just did not capture any eDNA. With my study results and previously published observations there is a total of 35 sightings of lamprey in Haldenvassdraget, with the majority of lamprey residing in the upper parts of the system (Figure 2 C and D).

## 4.3 Conclusions

eDNA methods can be better than traditional sampling according to some studies (Li et al., 2021; McColl-Gausden et al., 2021). It is however important to view eDNA results to be viewed in light of local knowledge, to hinder that wrong species being identified (Gargan et al., 2022). Even with the shortcoming of eDNA-metabarcoding, the consensus is that it is a good monitoring tool (Fujii et al., 2019; Gargan et al., 2022; Hänfling et al., 2016),

Both qPCR and DNA-metabarcoding worked as a tool to describe the distribution of river and brook lamprey in Haldenvassdraget. However, uncertainties of qPCR analyses failing or not, leads to inconclusive comparisons of detection probability, with qPCR potentially being statistically significantly more sensitive than DNA-metabarcoding.

My study suggests that DNA-metabarcoding is an acceptable monitoring tool for lamprey distribution, given the benefits it gives of a broader understanding of local biodiversity, and cost effectiveness as compared to qPCR.

## 5 Reference

- Agersnap, S., Larsen, W. B., Knudsen, S. W., Strand, D., Thomsen, P. F., Hesselsøe, M., Mortensen, P. B., Vrålstad, T., & Møller, P. R. (2017). Monitoring of noble, signal and narrow-clawed crayfish using environmental DNA from freshwater samples. *PLoS One*, *12*(6), e0179261.
- Ahmed, S. F., Kumar, P. S., Kabir, M., Zuhara, F. T., Mehjabin, A., Tasannum, N., Hoang, A. T., Kabir, Z., & Mofijur, M. (2022). Threats, challenges and sustainable conservation strategies for freshwater biodiversity. *Environmental Research*, 214, 113808.
- Alberdi, A., Aizpurua, O., Gilbert, M. T. P., & Bohmann, K. (2018). Scrutinizing key steps for reliable metabarcoding of environmental samples. *Methods in Ecology and Evolution*, 9(1), 134-147.
- Balasingham, K. D., Walter, R. P., & Heath, D. D. (2017). Residual eDNA detection sensitivity assessed by quantitative real-time PCR in a river ecosystem. *Molecular Ecology Resources*, 17(3), 523-532.
- Bessey, C., Jarman, S. N., Berry, O., Olsen, Y. S., Bunce, M., Simpson, T., Power, M., McLaughlin, J., Edgar, G. J., & Keesing, J. (2020). Maximizing fish detection with eDNA metabarcoding. *Environmental DNA*, 2(4), 493-504.
- Bylemans, J., Gleeson, D. M., Duncan, R. P., Hardy, C. M., & Furlan, E. M. (2019). A performance evaluation of targeted eDNA and eDNA metabarcoding analyses for freshwater fishes. *Environmental DNA*, *I*(4), 402-414.
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., & Holmes, S. P. (2016). DADA2: High-resolution sample inference from Illumina amplicon data. *Nature methods*, *13*(7), 581-583.
- Cowie, R. H., Bouchet, P., & Fontaine, B. (2022). The Sixth Mass Extinction: fact, fiction or speculation? *Biological Reviews*, 97(2), 640-663.
- Degerman, E. (2009). Fisk, fiske och miljö i de fyra stora sjöarna från istid till nutid. In: Fiskeriverket; Naturvårdsverket.
- Deiner, K., & Altermatt, F. (2014). Transport distance of invertebrate environmental DNA in a natural river. *PLoS One*, 9(2), e88786.
- Docker, M. F. (2014). Lampreys: Biology, Conservation and Control: Volume 1. In (Vol. 37): Springer.
- Eichmiller, J. J., Best, S. E., & Sorensen, P. W. (2016). Effects of temperature and trophic state on degradation of environmental DNA in lake water. *Environmental Science & Technology*, 50(4), 1859-1867.
- Elbrecht, V., & Leese, F. (2015). Can DNA-based ecosystem assessments quantify species abundance? Testing primer bias and biomass—sequence relationships with an innovative metabarcoding protocol. *PLoS One*, 10(7), e0130324.
- Fossøy, F., Thaulow, J., d'Auriac, M. A., Brandsegg, H., Sivertsgård, R., Mo, T. A., Sandlund, O. T., & Hesthagen, T. (2018). Bruk av miljø-DNA som supplerende verktøy for overvåkning og kartlegging av fremmed ferskvannsfisk.
- Fujii, K., Doi, H., Matsuoka, S., Nagano, M., Sato, H., & Yamanaka, H. (2019). Environmental DNA metabarcoding for fish community analysis in backwater lakes: A comparison of capture methods. *PLoS One*, *14*(1), e0210357.
- Gargan, L. M., Brooks, P. R., Vye, S. R., Ironside, J. E., Jenkins, S. R., Crowe, T. P., & Carlsson, J. (2022). The use of environmental DNA metabarcoding and quantitative PCR for molecular detection of marine invasive non-native species associated with artificial structures. *Biological Invasions*, 24(3), 635-648.

- Goodwin, C., Griffiths, D., Dick, J., & Elwood, R. (2006). A freshwater-feeding Lampetra fluviatilis L. population in Lough Neagh, Northern Ireland. *Journal of Fish Biology*, 68(2), 628-633.
- Greipsland, I., Bøe, F. N., & Turtumøygard, S. (2018). Overvåking av vannkvalitet i Haldenvassdraget 2017/2018. *NIBIO Rapport*.
- Harper, L. R., Lawson Handley, L., Hahn, C., Boonham, N., Rees, H. C., Gough, K. C., Lewis, E., Adams, I. P., Brotherton, P., & Phillips, S. (2018). Needle in a haystack? A comparison of eDNA metabarcoding and targeted qPCR for detection of the great crested newt (Triturus cristatus). *Ecology and evolution*, 8(12), 6330-6341.
- Hesthagen, T., Vøllestad, A., Brabrand, Å., Enerud, J., Fjellheim, A., Larsen, B. M., Sandlund, O. T., Sandaas, K., Schartum, E., & Spikkeland, I. (2021). Utbredelse av bekkeniøye i norske vassdrag [Popular science article]. *Fauna*, 74, 48-66.
- Huitfeldt-Kaas, H. J. (1918). Ferskvandsfiskenes utbredelse og innvandring i Norge medet tillceg om krebsen. Centraltrykkeriet, Kristiania. In: Norwegian In.
- Hänfling, B., Lawson Handley, L., Read, D. S., Hahn, C., Li, J., Nichols, P., Blackman, R. C., Oliver, A., & Winfield, I. J. (2016). Environmental DNA metabarcoding of lake fish communities reflects long-term data from established survey methods. *Molecular Ecology*, 25(13), 3101-3119.
- Jane, S. F., Wilcox, T. M., McKelvey, K. S., Young, M. K., Schwartz, M. K., Lowe, W. H., Letcher, B. H., & Whiteley, A. R. (2015). Distance, flow and PCR inhibition: e DNA dynamics in two headwater streams. *Molecular Ecology Resources*, *15*(1), 216-227.
- Johanson, A. (2002). *Haldenvassdraget*. Haldenvassdragets brukseierforening. <a href="https://doi.org/oai:nb.bibsys.no:990229664714702202">https://doi.org/oai:nb.bibsys.no:990229664714702202</a> URN:NBN:no-nb\_digibok\_2009050404015
- Johnsen, S. I., & Vrålstad, T. (2009). Signalkreps og krepsepest i Haldenvassdraget Forslag til tiltaksplan.
- Kelly, R. P., Shelton, A. O., & Gallego, R. (2019). Understanding PCR processes to draw meaningful conclusions from environmental DNA studies. *Scientific reports*, 9(1), 1-14
- Kuznetsova, A., Brockhoff, P. B., & Christensen, R. H. (2017). lmerTest package: tests in linear mixed effects models. *Journal of statistical software*, 82, 1-26.
- Lacoursière-Roussel, A., Rosabal, M., & Bernatchez, L. (2016). Estimating fish abundance and biomass from eDNA concentrations: variability among capture methods and environmental conditions. *Molecular Ecology Resources*, *16*(6), 1401-1414. https://doi.org/https://doi.org/10.1111/1755-0998.12522
- Lecaudey, L. A., Schletterer, M., Kuzovlev, V. V., Hahn, C., & Weiss, S. J. (2019). Fish diversity assessment in the headwaters of the Volga River using environmental DNA metabarcoding. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 29(10), 1785-1800.
- Li, W., Hou, X., Xu, C., Qin, M., Wang, S., Wei, L., Wang, Y., Liu, X., & Li, Y. (2021). Validating eDNA measurements of the richness and abundance of anurans at a large scale. *Journal of Animal Ecology*, 90(6), 1466-1479.
- McColl-Gausden, E. F., Weeks, A. R., Coleman, R. A., Robinson, K. L., Song, S., Raadik, T. A., & Tingley, R. (2021). Multispecies models reveal that eDNA metabarcoding is more sensitive than backpack electrofishing for conducting fish surveys in freshwater streams. *Molecular Ecology*, 30(13), 3111-3126.
- Miya, M., Gotoh, R. O., & Sado, T. (2020). MiFish metabarcoding: a high-throughput approach for simultaneous detection of multiple fish species from environmental DNA and other samples. *Fisheries Science*, 86(6), 939-970.

- Miya, M., Sato, Y., Fukunaga, T., Sado, T., Poulsen, J., Sato, K., Minamoto, T., Yamamoto, S., Yamanaka, H., & Araki, H. (2015). MiFish, a set of universal PCR primers for metabarcoding environmental DNA from fishes: detection of more than 230 subtropical marine species. *Royal Society open science*, *2*(7), 150088.
- Muha, T. P., Robinson, C. V., Garcia de Leaniz, C., & Consuegra, S. (2019). An optimised eDNA protocol for detecting fish in lentic and lotic freshwaters using a small water volume. *PLoS One*, *14*(7), e0219218.
- Nichols, R. V., Vollmers, C., Newsom, L. A., Wang, Y., Heintzman, P. D., Leighton, M., Green, R. E., & Shapiro, B. (2018). Minimizing polymerase biases in metabarcoding. *Molecular Ecology Resources*, 18(5), 927-939.
- Pethon, P. (2005). Aschehougs store fiskebok. Norges fisker i farger. Aschehoug.
- Piñol, J., Senar, M. A., & Symondson, W. O. (2019). The choice of universal primers and the characteristics of the species mixture determine when DNA metabarcoding can be quantitative. *Molecular Ecology*, 28(2), 407-419.
- Pont, D., Rocle, M., Valentini, A., Civade, R., Jean, P., Maire, A., Roset, N., Schabuss, M., Zornig, H., & Dejean, T. (2018). Environmental DNA reveals quantitative patterns of fish biodiversity in large rivers despite its downstream transportation. *Scientific reports*, 8(1), 1-13.
- Rees, H. C., Maddison, B. C., Middleditch, D. J., Patmore, J. R., & Gough, K. C. (2014). The detection of aquatic animal species using environmental DNA–a review of eDNA as a survey tool in ecology. *Journal of Applied Ecology*, 51(5), 1450-1459.
- Rourke, M. L., Fowler, A. M., Hughes, J. M., Broadhurst, M. K., DiBattista, J. D., Fielder, S., Wilkes Walburn, J., & Furlan, E. M. (2022). Environmental DNA (eDNA) as a tool for assessing fish biomass: A review of approaches and future considerations for resource surveys. *Environmental DNA*, 4(1), 9-33.
- Roussel, J.-M., Paillisson, J.-M., Treguier, A., & Petit, E. (2015). The downside of eDNA as a survey tool in water bodies. *Journal of Applied Ecology*, 823-826.
- Ruppert, K. M., Kline, R. J., & Rahman, M. S. (2019). Past, present, and future perspectives of environmental DNA (eDNA) metabarcoding: A systematic review in methods, monitoring, and applications of global eDNA. *Global Ecology and Conservation*, 17, e00547.
- Schartum, E., & Kraabøl, M. (2013). *Undersøkelser av bekkeniøye i Bandakdeltaet og Tokkeåi, Telemark Resultater fra undersøkel*□*sene i 2012 og 2013*.
- Schenekar, T., Schletterer, M., Lecaudey, L. A., & Weiss, S. J. (2020). Reference databases, primer choice, and assay sensitivity for environmental metabarcoding: Lessons learnt from a re-evaluation of an eDNA fish assessment in the Volga headwaters. *River Research and Applications*, 36(7), 1004-1013.
- Schrader, C., Schielke, A., Ellerbroek, L., & Johne, R. (2012). PCR inhibitors—occurrence, properties and removal. *Journal of applied microbiology*, *113*(5), 1014-1026.
- Selbekk, L. K., Bergerud, J., Kollerud, E., & Krøgenes, N. A. (2020). Kartlegging av ål i Haldenvassdraget. Prøvefiske med åleruser i Bjørkelangen, Rødenessjøen, Aremarksjøen og Femsjøen.
- Spikkeland, I. (2014). Biologisk mangfold i Haldenvassdraget. Om planter og dyr knyttet til vann i vassdragets nedbørfelt. *Rapport 1/2014*.
- Spikkeland, I., & Kasbo, R. (2014). Bekkeniøye Lampetra planeri og elveniøye Lampetra fluviatilis i Haldenvassdraget. *Natur i Østfold*, *33*, 3-7.
- Strand, D. A., Johnsen, S. I., Rusch, J. C., Agersnap, S., Larsen, W. B., Knudsen, S. W., Møller, P. R., & Vrålstad, T. (2019). Monitoring a Norwegian freshwater crayfish tragedy: eDNA snapshots of invasion, infection and extinction. *Journal of Applied Ecology*, 56(7), 1661-1673.

- Team, R. C. (2021). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. https://www.R-project.org/.
- Thomsen, P. F., Kielgast, J., Iversen, L. L., Wiuf, C., Rasmussen, M., Gilbert, M. T. P., Orlando, L., & Willerslev, E. (2012). Monitoring endangered freshwater biodiversity using environmental DNA. *Molecular Ecology*, 21(11), 2565-2573.
- Tremblay, J., Singh, K., Fern, A., Kirton, E. S., He, S., Woyke, T., Lee, J., Chen, F., Dangl, J. L., & Tringe, S. G. (2015). Primer and platform effects on 16S rRNA tag sequencing. *Frontiers in microbiology*, *6*, 771.
- Turner, C. R., Barnes, M. A., Xu, C. C., Jones, S. E., Jerde, C. L., & Lodge, D. M. (2014). Particle size distribution and optimal capture of aqueous macrobial eDNA. *Methods in Ecology and Evolution*, *5*(7), 676-684.
- Valentini, A., Taberlet, P., Miaud, C., Civade, R., Herder, J., Thomsen, P. F., Bellemain, E., Besnard, A., Coissac, E., & Boyer, F. (2016). Next-generation monitoring of aquatic biodiversity using environmental DNA metabarcoding. *Molecular Ecology*, 25(4), 929-942.
- Wacker, S., Fossøy, F., Larsen, B. M., Brandsegg, H., Sivertsgård, R., & Karlsson, S. (2019). Downstream transport and seasonal variation in freshwater pearl mussel (Margaritifera margaritifera) eDNA concentration. *Environmental DNA*, *I*(1), 64-73. https://doi.org/https://doi.org/10.1002/edn3.10
- Wilcox, T. M., McKelvey, K. S., Young, M. K., Sepulveda, A. J., Shepard, B. B., Jane, S. F., Whiteley, A. R., Lowe, W. H., & Schwartz, M. K. (2016). Understanding environmental DNA detection probabilities: A case study using a stream-dwelling char Salvelinus fontinalis. *Biological Conservation*, 194, 209-216.
- Wood, S. A., Pochon, X., Laroche, O., von Ammon, U., Adamson, J., & Zaiko, A. (2019). A comparison of droplet digital polymerase chain reaction (PCR), quantitative PCR and metabarcoding for species-specific detection in environmental DNA. *Molecular Ecology Resources*, 19(6), 1407-1419.
- Wrona, F. J., Prowse, T. D., Reist, J. D., Hobbie, J. E., Lévesque, L. M., & Vincent, W. F. (2006). Climate change effects on aquatic biota, ecosystem structure and function. *AMBIO: A Journal of the Human Environment*, 35(7), 359-369.
- Yu, Z., Ito, S.-i., Wong, M. K.-S., Yoshizawa, S., Inoue, J., Itoh, S., Yukami, R., Ishikawa, K., Guo, C., & Ijichi, M. (2022). Comparison of species-specific qPCR and metabarcoding methods to detect small pelagic fish distribution from open ocean environmental DNA. *PLoS One*, 17(9), e0273670.
- Zancolli, G., Foote, A., Seymour, M., & Creer, S. (2018). Assessing lamprey populations in Scottish rivers using eDNA: proof of concept (1783914572).

## 6 Appendixes

Appendix table 1 List of study sites in Haldenvassdraget. The volume of water (L) sampled for replicate A and B, and water temperature ( $^{\circ}$ C) is given

Sample ID	Date	Time	Locality	Latitude	Longitude	Water volume	Water temperature
01A	07.06.2021	09:30	Ulviksjøen	59.966797	11.495812	2	20
01B	07.06.2021	09:30	Ulviksjøen	59.966797	11.495812	2.5	20
02A	07.06.2021	10:50	Eidsbekken	59.884123	11.580333	2	20.8
02B	07.06.2021	10:50	Eidsbekken	59.884123	11.580333	3	20.8
03A	07.06.2021	11:30	Lierelva	59.886227	11.575236	1	20.5
03B	07.06.2021	11:30	Lierelva	59.886227	11.575236	1	20.5
04A	07.06.2021	12:10	Elv fra Snartjern	59.840777	11.538131	1	16.4
04B	07.06.2021	12:10	Elv fra Snartjern	59.840777	11.538131	1	16.4
05A	07.06.2021	13:25	Malttjernbekken	59.836745	11.479805	1.5	15.2
05B	07.06.2021	13:25	Malttjernbekken	59.836745	11.479805	2	15.2
06A	07.06.2021	14:00	Dalselva	59.727686	11.404894	1.5	15
06B	07.06.2021	14:00	Dalselva	59.727686	11.404894	1	15
07A	07.06.2021	14:45	Hemneselva	59.725505	11.420211	3	21.5
07B	07.06.2021	14:45	Hemneselva	59.725505	11.420211	2	21.5
08A	07.06.2021	15:40	Lower Hafsteinselva	59.722643	11.466323	2	21
08B	07.06.2021	15:40	Lower Hafsteinselva	59.722643	11.466323	1	21
09A	07.06.2021	16:15	Hølansaelva	59.720461	11.492098	1.5	20.3
09B	07.06.2021	16:15	Hølansaelva	59.720461	11.492098	1	20.3
10A	07.06.2021	16:50	Upper Hafsteinselva	59.747636	11.435987	1	17
10B	07.06.2021	16:50	Upper Hafsteinselva	59.747636	11.435987	1	17
11A	08.06.2021	09:50	Setta	59.84663	11.675412	1	16.5
11B	08.06.2021	09:50	Setta	59.84663	11.675412	1	16.5
12A	08.06.2021	10:45	Langebruslora	59.787913	11.726863	2	19
12B	08.06.2021	10:45	Langebruslora	59.787913	11.726863	2	19
13A	08.06.2021	11:45	Langtjen	59.810876	11.61592	5	19
13B	08.06.2021	11:45	Langtjen	59.810876	11.61592	4	19
14A	08.06.2021	13:40	Upper Mjerma	59.713321	11.555722	4	18.2
14B	08.06.2021	13:40	Upper Mjerma	59.713321	11.555722	4.00	18.20
15A	08.06.2021	14:30	Lower Mjerma	59.694242	11.528943	4	18.2
15B	08.06.2021	14:30	Lower Mjerma	59.694242	11.528943	3	18.2
16A	08.06.2021	16:45	Nautebrofoss	59.618259	11.596371	1.25	17.9
16B	08.06.2021	16:45	Nautebrofoss	59.618259	11.596371	2	17.9
17A	09.06.2021	13:00	Østenbyelva	59.533827	11.60172	1.5	19.2
17B	09.06.2021	13:00	Østenbyelva	59.533827	11.60172	2	19.2
18A	09.06.2021	13:45	Taraldruelva	59.581574	11.565786	1.5	15.6
18B	09.06.2021	13:45	Taraldruelva	59.581574	11.565786	2	15.6

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19A	09.06.2021	14:30	Åsebyelva	59.600175	11.61974	1.5	17
19B	09.06.2021	14:30	Åsebyelva	59.600175	11.61974	1	17
20A	09.06.2021	15:15	Langnes	59.546237	11.626861	2	17.2
20B	09.06.2021	15:15	Langnes	59.546237	11.626861	2.5	17.2
21A	09.06.2021	16:00	Engerelva	59.491057	11.667821	1.5	17.8
21B	09.06.2021	16:00	Engerelva	59.491057	11.667821	1.5	17.8
22A	10.06.2021	08:30	Ørje	59.472415	11.654516	2.5	13.7
22B	10.06.2021	08:30	Ørje	59.472415	11.654516	3	13.7
23A	10.06.2021	09:20	Bøenselva	59.461618	11.660805	2	18.4
23B	10.06.2021	09:20	Bøenselva	59.461618	11.660805	2	18.4
24A	10.06.2021	09:50	Gunnengbekken	59.366216	11.689881	1	14.3
24B	10.06.2021	09:50	Gunnengbekken	59.366216	11.689881	1	14.3
25A	10.06.2021	10:40	Skinnarbutjenn	59.351489	11.733146	3	19.2
25B	10.06.2021	10:40	Skinnarbutjenn	59.351489	11.733146	2.5	19.2
26A	10.06.2021	11:20	Langetjernelva	59.329994	11.685982	2	18.3
26B	10.06.2021	11:20	Langetjernelva	59.329994	11.685982	2	18.3
27A	10.06.2021	12:00	Fangebekken	59.288988	11.678935	1	14.9
27B	10.06.2021	12:00	Fangebekken	59.288988	11.678935	1	14.9
28A	10.06.2021	12:40	Fossby	59.220603	11.6962	1	16.3
28B	10.06.2021	12:40	Fossby	59.220603	11.6962	1	16.3
29A	10.06.2021	13:20	Skolleborg	59.232771	11.671846	1.5	18.7
29B	10.06.2021	13:20	Skolleborg	59.232771	11.671846	1	18.7
30A	10.06.2021	14:00	Verksbrua	59.25494	11.64334	2.5	20.1
30B	10.06.2021	14:00	Verksbrua	59.25494	11.64334	2.5	20.1
31A	10.06.2021	14:30	Tenebekken	59.280254	11.65288	1	15.3
31B	10.06.2021	14:30	Tenebekken	59.280254	11.65288	1	15.3
32A	10.06.2021	15:20	Strømsfoss	59.301061	11.658639	3	16.6
32B	10.06.2021	15:20	Strømsfoss	59.301061	11.658639	3	16.6
33A	11.06.2021	08:30	Lielva	59.461153	11.644340	1.5	18.8
33B	11.06.2021	08:30	Lielva	59.461153	11.644340	1.5	18.8
34A	11.06.2021	09:10	Valbyelva	59.381098	11.642453	1	14.5
34B	11.06.2021	09:10	Valbyelva	59.381098	11.642453	1	14.5
35A	11.06.2021	10:15	Holmegilelva	59.142629	11.739045	4	20.8
35B	11.06.2021	10:15	Holmegilelva	59.142629	11.739045	4	20.8
36A	11.06.2021	11:20	Kverntjern (Remne)	59.140828	11.660145	2	20.5
36B	11.06.2021	11:20	Kverntjern (Remne)	59.140828	11.660145	2	20.5
37A	11.06.2021	12:10	Stenselva	59.14554	11.638203	3.5	15.5
37B	11.06.2021	12:10	Stenselva	59.14554	11.638203	3	15.5
38A	11.06.2021	13:00	Ganerødelva	59.144157	11.534509	2.33	20.4
38B	11.06.2021	13:00	Ganerødelva	59.144157	11.534509	2.2	20.4
39A	11.06.2021	14:10	Mellebyelva	59.180468	11.689811	1	20.1
39B	11.06.2021	14:10	Mellebyelva	59.180468	11.689811	1	20.1
40A	13.06.2021	14:10	Rødselva	59.171888	11.545509	1	20.2
40B	13.06.2021	14:10	Rødselva	59.171888	11.545509	1	20.2
41A	13.06.2021	14:50	Rjørelva	59.183022	11.51762	1.5	21.3
41B	13.06.2021	14:50	Rjørelva	59.183022	11.51762	1.5	21.3
4TD	13.00.2021	14.50	njørerva	33.103022	11.51/02	1.5	21.3

42A	13.06.2021	15:30	Asakbekken	59.138606	11.451391	1	13.7
42B	13.06.2021	15:30	Asakbekken	59.138606	11.451391	1	13.7
43A	13.06.2021	16:30	Tistedalsfoss	59.130336	11.437732	4	15.3
43B	13.06.2021	16:30	Tistedalsfoss	59.130336	11.437732	4	15.3
44A	13.06.2021	18:00	Skåningsfoss	59.125933	11.416768	5	21
44B	13.06.2021	18:00	Skåningsfoss	59.125933	11.416768	5	21
45A	13.06.2021	20:00	Porsnes sluser	59.123438	11.39958	4	20.5
45B	13.06.2021	20:00	Porsnes sluser	59.123438	11.39958	4	20.5

Appendix table 2 Optical density (OD) showing DNA-quantity and quality of the DNA-extracts.

Camarda ID	00	200 200	200 220
Sample ID	OD	260-280	260-230
01A	67.17	1.57	1.16
01B	84.79	1.63	1.15
02A	154.30	1.62	0.91
02B	168.60	1.67	0.94
03A	354.90	1.73	1.26
03B	345.80	1.70	1.19
04A	177.40	1.59	0.93
04B	161.10	1.56	0.87
05A	92.13	1.65	0.95
05B	115.80	1.69	1.09
06A	308.00	1.53	0.90
06B	250.90	1.70	1.05
07A	266.60	1.94	1.73
07B	193.20	1.93	1.63
08A	268.40	1.61	1.02
08B	275.70	1.83	1.38
09A	327.50	1.74	1.23
09B	232.60	1.84	1.47
10A	312.80	1.63	1.00
10B	328.50	1.68	1.12
11A	117.90	1.68	1.13
11B	45.57	1.51	1.07
12A	96.90	1.74	1.33
12B	90.04	1.70	1.17
13A	242.20	1.76	1.18
13B	84.02	1.60	1.10
14A	224.20	1.63	0.99
14B	65.22	1.70	0.91
15A	283.30	1.72	1.18
15B	133.50	1.67	1.07
16A	171.50	1.60	1.00
16B	141.90	1.61	0.99
17A	105.30	1.78	1.43

17B	187.40	1.87	1.50
18A	895.30	1.90	1.31
18B	494.80	1.78	1.31
19A	23.79	1.87	6.00
19B	67.37	1.70	1.57
20A	250.40	1.70	1.13
20B	214.60	1.70	1.11
21A	102.00	1.65	1.17
21B	135.70	1.82	1.43
22A	418.20	1.72	1.29
22B	421.50	1.79	1.39
23A	423.20	1.60	0.94
23B	387.40	1.65	0.92
24A	284.00	1.86	1.52
24B	247.80	1.78	1.32
25A	342.30	1.80	1.38
25B	222.40	1.71	1.15
26A	135.80	1.78	1.38
26B	129.90	1.86	1.48
27A	102.80	1.68	1.35
27B	96.26	1.62	1.28
28A	177.10	1.60	0.84
28B	58.52	1.44	0.64
29A	129.9	1.72	1.17
29B	91.32	1.69	1.17
30A	43.7	1.47	0.7
30B	45.2	1.65	1.32
31A	429.6	1.36	0.86
31B	160	1.79	1.23
32A	548.5	1.71	1.37
32B	499.1	1.76	1.49
33A	78.36	1.73	1.38
33B	125.7	1.78	1.41
34A	277.3	1.82	1.36
34B	156.4	1.8	1.35
35A	70.06	1.4	0.58
35B	76.13	1.45	0.57
36A	173.3	1.62	0.96
36B	97.32	1.51	0.86
37A	406.4	1.74	1.33
37B	240.1	1.89	1.57
38A	151.7	1.73	1.22
38B	101.9	1.76	1.59
39A	316.2	1.63	0.91
39B	153.3	1.39	0.51
40A	77.62	1.72	1.27
		_	

40B	238.3	1.9	1.75
41A	543.8	1.65	1.04
41B	687.4	1.71	0.96
42A	367.4	1.96	1.68
42B	298.9	1.97	1.73
43A	173.6	1.78	1.35
43B	375.3	1.46	0.97
44A	242.7	1.55	0.92
44B	566.9	1.69	1.41
45A	394.5	1.85	1.57
45B	544.1	1.71	1.4

Appendix table 3 No of reads in the different filtering steps as part of the bioinformatic pipeline.

PrøvelD	raw	input	filtered	denoisedF	denoisedR	merged	nonchim	pct_retained
01A	1455086	1062405	948304	946120	945593	933420	921167	63.31
01B	1323113	935574	856976	854968	854182	843781	832564	62.92
02A	1657246	1294966	1150916	1148201	1148250	1129475	1099906	66.37
02B	1583681	1239705	1114008	1111423	1110675	1087359	1050906	66.36
03A	1158269	824013	743885	742136	741844	732210	722124	62.35
03B	1235933	878380	777985	775945	775860	765656	753274	60.95
04A	1646107	1380993	1261281	1256340	1255571	1221950	1183165	71.88
04B	1580171	1353233	1227704	1222223	1221426	1183311	1136069	71.90
05A	953837	691860	588129	586627	586214	577255	568899	59.64
05B	1156764	876671	793951	792015	791769	780411	767953	66.39
06A	790254	481334	434460	434177	433991	431932	426385	53.96
06B	867029	380806	338488	338200	338224	334813	332007	38.29
07A	1009808	683317	505775	505035	505062	501577	496395	49.16
07B	807399	479007	434217	433858	433892	431838	429060	53.14
A80	429139	263564	239376	239135	239014	235784	231809	54.02
08B	374327	213285	194728	194211	194017	191270	187355	50.05
09A	1308500	918702	785450	783602	783243	774497	764131	58.40
09B	820822	441237	391429	391044	391066	389560	386818	47.13
10A	1379215	1034237	936521	934166	934210	913186	898145	65.12
10B	971990	738383	653938	652063	651668	636529	624973	64.30
11A	1227277	845897	761377	759424	759014	743635	729806	59.47
11B	1193528	225942	177654	177533	177543	177015	176399	14.78
12A	1182251	748677	678988	677972	677351	670813	665332	56.28
12B	1135518	704805	626534	625411	625283	618903	611714	53.87
13A	1139099	742628	446291	445760	444883	441420	436388	38.31
13B	1299729	841699	756843	755174	754793	746359	739882	56.93
14A	1270575	487565	401367	401123	401227	399924	399600	31.45

14B	496158	370678	307604	307571	307534	305114	304399	61.35
15A	998805	733936	486371	485622	485664	481865	476432	47.70
15B	537641	382722	338714	338054	337795	334494	331098	61.58
16A	398782	264403	224284	223832	223847	221247	218321	54.75
16B	545654	418196	377990	376844	376760	370862	365093	66.91
17A	904553	536732	379727	379624	379553	378345	377946	41.78
17B	848947	424423	376652	376450	376329	373886	372382	43.86
18A	1848867	1439197	1293461	1290026	1290000	1265128	1234862	66.79
18B	2186670	1752046	1613300	1608670	1608115	1568269	1519424	69.49
19A	1494406	1195186	1088603	1083408	1083047	1049092	1016677	68.03
19B	1225045	951403	793783	791652	790931	774680	758840	61.94
20A	1123666	917989	777969	774809	773949	754933	741848	66.02
20B	1137157	905002	832951	829918	829179	811403	795849	69.99
21A	1531607	1191316	1098824	1094632	1094704	1073616	1053905	68.81
21B	2261556	1814918	1538957	1535452	1534625	1508322	1483242	65.59
22A	774687	563375	508850	507789	507515	502850	496382	64.08
22B	939129	697651	608640	607622	607253	601536	594419	63.29
23A	578616	417497	372675	371593	371550	367681	361511	62.48
23B	396444	303459	265098	264033	263961	260242	257687	65.00
24A	1305015	1028230	903256	899828	899393	876781	856967	65.67
24B	1756760	1371703	1260156	1256162	1255146	1223837	1190277	67.75
25A	1218850	857455	684772	683594	683533	677963	669556	54.93
25B	1242227	855140	778782	776714	776227	766422	754691	60.75
26A	1237381	735624	480504	480228	480196	477621	470957	38.06
26B	1447843	1135961	1023593	1019834	1019139	998772	981783	67.81
27A	1130362	943496	840522	837904	837069	820089	803981	71.13
27B	1070393	911079	726092	723710	723721	708962	696787	65.10
28A	973822	824078	640042	638869	638596	629680	618090	63.47
28B	483830	225461	37467	37440	37383	37261	37256	7.70
29A	795269	598800	446351	445754	445700	439883	432938	54.44
29B	1109002	838214	757921	756784	755896	747483	738364	66.58
30A	686691	494429	229701	229655	229609	229161	228859	33.33
30B	2238568	1606631	1448081	1444321	1443283	1418370	1387461	61.98
31A	382723	290621	106369	106254	106321	106062	106060	27.71
31B	817229	632536	551571	549445	549028	536695	525289	64.28
32A	1240336	868960	776394	775185	774794	768794	757618	61.08
32B	1007850	720262	662772	661753	661579	657486	652457	64.74
33A	1209018	952436	869812	866422	866251	845386	827429	68.44
33B	1205502	760866	683414	682511	682696	676816	658954	54.66
34A	1634446	1319000	1110080	1107260	1106883	1086620	1064472	65.13
34B	1171357	750115	497545	497023	496976	494541	491411	41.95
35A	38970	33375	10632	10609	10621	10552	10551	27.07
35B	416987	172526	118477	118462	118433	118153	118090	28.32
36A	765411	526251	480079	479938	479933	478764	471625	61.62
36B	62624	59253	5	479938	479933	2	4/1023	0.00
37A	1544958	1027355	931407	930223	930142	923167	908661	58.81
3/A	1344938	102/333	331407	330223	550142	32310/	100001	38.81

37B	766127	538888	212066	211860	211843	211003	209867	27.39
38A	793794	584364	523985	523753	523254	519814	518056	65.26
38B	736857	580872	516957	516259	516227	512539	508283	68.98
39A	914209	673467	601978	601683	601638	597243	592529	64.81
39B	216904	122869	59517	59497	59486	59383	59377	27.37
40A	168033	100918	41722	41652	41654	41543	41535	24.72
40B	1303004	869905	794138	792251	791838	780668	767773	58.92
41A	1002971	675493	614587	614256	614357	611668	605568	60.38
41B	909027	688181	511351	511059	510798	508640	503175	55.35
42A	1541405	1192740	1095746	1092988	1091944	1068436	1042102	67.61
42B	1472582	1142650	1037900	1033887	1032775	998808	969002	65.80
43A	937270	576230	515045	513988	513602	508837	502874	53.65
43B	1402609	749325	682932	681878	681727	675816	670631	47.81
44A	1038517	426209	369992	369682	369601	367050	365709	35.21
44B	1010522	579909	514339	513362	512757	509963	505451	50.02
45A	996399	612092	559950	558916	558441	554389	550469	55.25
45B	431615	254855	228299	227733	227595	226173	224766	52.08
NEG_150721	25190	9898	45	45	43	43	43	0.17
NEG_220721	123358	105000	43	43	43	43	43	0.03
NEG1	56199	36257	7901	7894	7894	7894	7841	13.95
NEG2	139959	112760	1818	1812	1812	1812	1809	1.29
NEG3	31462	17151	0	NA	NA	NA	NA	NA

Appendix table 4 DNA-metabarcoding results. All fish species found in the different samling location. With the species eDNA abundance.

